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**Effect of seawater-freshwater cross-transplantations on viral dynamics and
bacterial diversity and production**

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Abstract

Dilution experiments were carried out to investigate the community composition and the metabolic response of seawater and freshwater bacteria to cross-transplantation, and the effects of non-indigenous bacterial hosts on viral dynamics. Changes in viral and bacterial abundance and production, as well as bacterial respiration, carbon demand and diversity were regularly monitored over a 6 day period. Bacterial production in the transplanted seawater and freshwater bacteria (SB-t and FB-t treatments) was stimulated up to 256% and 221%, respectively, compared to controls. The stimulation of bacterial production and carbon demand was accompanied by a decrease in bacterial richness. Net viral production was stimulated by 81% in SB-t and repressed by 75% in FB-t. Transplantation increased the virus-induced mortality of marine bacteria but decreased it for freshwater bacteria. These results suggest that (1) marine bacteria can readily oxidize freshwater dissolved organic matter (DOM), and (2) that freshwater viruses might be able to infect marine hosts, thus highlighting their potential role in fueling bacterial growth under resource stress or nutrient-depleted conditions.

Introduction

Substrate availability has been extensively studied as a potentially important factor controlling prokaryotic activity. It is well known that the quality and quantity of inorganic and organic nutrients can exert significant control on prokaryotic structure and function (e.g. Azam and Malfatti (2007)). In coastal areas, and particularly in estuarine systems, large variations in nutrient and organic carbon concentration can occur over relatively small distances. For example, in Randers Fjord, a small estuary (27 km long) in Northern Denmark, concentrations in inorganic and organic nitrogen increase by 1 to 3 orders of magnitude between the seawater and freshwater end-members, respectively (Veuger et al. 2004). Similarly, dissolved organic carbon concentrations also vary along estuarine gradients with generally higher concentrations in the freshwater end-members and lower concentrations in the seawater end-members (Fisher et al. 1998, Abril et al. 2002). Superimposed upon these gradients, biological processes both modify and are modified by geochemical processes.

The dynamic nature of estuaries means that solutes and organic matter from the freshwater and seawater end-members become mixed along the estuarine gradient. The manner in which these inputs mix is related to the relative proportions of each input, as well as to other physical properties (tidal regime, winds, etc). Therefore, communities of bacteria and other planktonic organisms are exposed to strong physicochemical gradients and constantly varying environmental conditions along the length of the estuary. Moreover, in estuaries where freshwater inputs are relatively low compared to that of seawater inputs, such as in the Scheldt estuary, bacteria and other organisms experience a situation where salinity changes are reduced but where relatively large changes in dissolved organic matter (DOM) and other solutes can occur. This results in the exposure of seawater bacteria to freshwater DOM. The reverse side of this, of course, is that in these systems with high seawater to freshwater inputs, freshwater bacteria will be subject to the dual problems of changing salinity and DOM through mixing with

seawater. Estuaries are therefore attractive systems to address the question of substrate availability as a controlling factor of prokaryotic activity (Jones et al. 2007).

Strong compositional and metabolic changes have been shown to occur in bacterial communities between the freshwater and saltwater portions of estuaries (Bouvier & del Giorgio 2002, Kirchman et al. 2004) and several authors have proposed the existence of a unique estuarine community that is found in the mixing zone (e.g. Crump et al. 2004). Indeed, the domination of a unique estuarine community vs. a mixing community in the middle estuary largely depends on the water residence time (Bouvier & del Giorgio 2002, Crump et al. 2004, Kan et al. 2006). However, regardless of water residence time, the instability of the environment is accompanied by changes in bacterial metabolism with bacterial cells more dedicated to physiological maintenance than cell production (del Giorgio & Bouvier 2002). Moreover, a recent meta-analysis of bacterial diversity data has shown that salinity plays an important role in structuring bacterial communities (Lozupone & Knight 2007) and it also seems to be a determining factor in carbon substrate utilisation in estuaries (Thottathil et al. 2008).

Although several studies have recently investigated bacterial responses to changes in the supply of organic matter in estuaries (Stepanauskas et al. 1999, Findlay et al. 2003, Langenheder et al. 2004), few have considered the impact of the associated variation of the coexisting viral communities, another key factor controlling bacterial dynamics. It is now well established that virioplankton exerts a “top-down” pressure and is a significant mortality agent for heterotrophic bacteria. Virioplankton also plays a significant role in shaping the composition and controlling the diversity of its hosts (Thingstad 2000, Wommack & Colwell 2000) and up to 50% of bacterial mortality can be due to viral activity (Fuhrman & Schwalbach 2003, Weinbauer 2004, Bouvier & del Giorgio 2007). Viral activity therefore has important effects on bacterial processes in aquatic systems. Over and above the direct effect of viruses on infected prokaryotic cells, viral lysis can enhance the activity of non-infected prokaryotic cells, which benefit from

the release of organic matter by cell lysis (Middelboe et al. 1996). Recent work also highlighted the complexity of the interactions between viral and bacterial activity and showed that viral lysis does not always result in increased bacterial growth efficiency (Bonilla-Findji et al. 2008). These authors reported that BR was stimulated (up to 113%) in the presence of active viruses whereas BP and BGE were reduced (up to 51%) and suggested that viruses enhance the role of bacteria as oxidizers of organic matter and as producers of CO₂.

Similar to that observed for bacteria, spatio-temporal changes in the viroplankton composition and structure have been observed in estuaries (Wommack et al. 1999) and different viral infection rates are known to occur along salinity gradients (Almeida et al. 2001). Burst size, and thus viral production, tend to increase as a function of increasing DOM and nutrient concentration (Bettarel et al. 2004, Parada et al. 2006). It is therefore probable that changes in DOM in terms of quality and quantity play a role in determining viral activity in estuarine systems.

The present study was carried out in the Scheldt Estuary (SW Netherlands and NE Belgium), one of the most eutrophic estuaries in Europe (Wollast 1988) in order to extend previous results on the functional response of a seawater bacterial community to freshwater dissolved organic matter (Rochelle-Newall et al. 2004). Two objectives were addressed: (1) to determine the structural and metabolic response of seawater and freshwater bacterial communities to cross-transplantation and (2) to investigate how viral dynamics change during transplantation relative to changes in DOM concentration and host diversity.

Methods

Experimental setup. In order to assess the response of bacteria and viruses to mixing along an estuarine salinity gradient, water samples were collected in April 2003 from the seawater and freshwater end-members of the Scheldt estuary (Fig. 1, Table 1). The response of seawater bacteria to freshwater DOM and viruses and, conversely, the response of freshwater bacteria to seawater DOM and viruses were investigated using dilution experiments. This transplant method was chosen to experimentally mimic the dynamic mixing of water masses and hence bacterial and viral communities and DOM along the estuarine salinity gradient of the Scheldt estuary.

Bacteria and viruses from the two end-members of the estuary were inoculated in filtered water from the other site in a 1:18 volume ratio (Fig. 2). At each station, 18 L of $< 0.2 \mu\text{m}$ filtrate water were distributed into two 20 L, acid washed and Milli-Q rinsed polycarbonate carboys and then inoculated with 1 L of either unfiltered seawater or unfiltered freshwater in order to obtain four treatments: transplanted seawater bacteria (SB-t), seawater control (SB-c); transplanted freshwater bacteria (FB-t) and freshwater control (FB-c). The salinity of the freshwater filtrate was adjusted to 30 with an artificial seawater (ASW; Guillard 1975) mix to minimize any effect of the salinity on the seawater bacterial community.

The four treatments were incubated for 6 d in the dark and at *in situ* temperature (7.6°C) and sub-samples were collected for bacteria and viral counts, bacterial production and diversity at six time points (0h; 17h; 43h; 91h; 139h; 157h). Samples for determination of dissolved organic carbon (DOC) were taken at the beginning and end of the experiment.

Sample analyses. Viral abundance was measured in duplicate, 1 ml samples fixed with glutaraldehyde (0.5% final concentration, EM-grade; Merck) for 30 min at 4°C in the dark, flash

frozen in liquid nitrogen and stored at -80°C pending analysis by flow cytometry as described by Brussaard (2004).

Viral counts were performed with the CellQuest-Pro software (Becton Dickinson and Company) after staining with SYBR GREEN I (Invitrogen: S-7567) that was pre-diluted in $0.2\ \mu\text{m}$ filtered, autoclaved Milli-Q (final dilution of 5×10^{-5} the commercial stock). Three viral groups were discriminated from scatter plots of side scatter (SSC) versus green fluorescence (FL1). These groups were labeled according to their increasing fluorescence signal: V1 (low), V2 (medium) and V3 (high). Previous comparisons (including the studied environment) of the total counts obtained by this method with viral abundances determined by microscopy were very similar and differences were generally less than 20% (O. Bonilla-Findji, unpublished). Viral production and decay was calculated from the net changes in viral abundance between each time point (Bratbak et al. 1990). The values obtained should therefore be considered as conservative estimations of viral production and decay.

Bacterial abundance was measured by a direct count method using epifluorescence microscopy and DAPI-stained samples (Porter & Feig 1980). Inspection of filters during enumeration did not reveal that flagellates were present in significant numbers. Bacterial production was estimated from ^3H -leucine incorporation in accordance with the method of Kirchman (1992) and as previously described elsewhere (Rochelle-Newall et al. 2004). Bacterial respiration was calculated based on standard dark-bottle O_2 consumption rates. Replicate BOD bottles were incubated in the dark following the JGOFS protocol (Knap et al. 1996) and at distinct time points, oxygen concentration was titrated using an automated Winkler titration technique with a potentiometric end-point detection (Anderson et al. 1992). The respiration rate at each time point T and its standard error were determined by regressing O_2 concentration against time during the intervals of incubation T to $T+I$.

Due to the difference in the time scales of the measurement of bacterial production (1h) and

bacterial respiration (17 to 48 h), bacterial carbon demand (= bacterial production + bacterial respiration) was calculated using bacterial respiration, expressed in carbon units assuming a respiratory quotient of 1. Mean bacterial production was determined during the same time interval as used to measure the rate of respiration using $((BP_{T1} + BP_{T2})/2)$, BP_{T1} and BP_{T2} are bacterial production at the start and end of the respiration measurement, respectively.

Samples (10ml) for determination of dissolved organic carbon were filtered through combusted (450°C, overnight) glass fiber filters (Whatman GF/F) and sealed in pre-combusted (450°C for 4h) glass ampoules after adding 12 µl of 85% v/v phosphoric acid (H_3PO_4). Samples were stored in the dark at 4°C pending analysis. DOC concentration was determined as previously described (Rochelle-Newall et al. 2007) by the high temperature combustion method using a Shimadzu TOC-5000 analyzer. Certified reference materials (Hansell Laboratory, University of Miami) were also used to assess the performance of the instrument on and between measurement days (Sharp 2002).

For extraction of DNA from prokaryotic cells 50 to 150 ml samples were recovered on a 0.2 µm pore-size polycarbonate filter (47 mm diameter; Whatman) and kept frozen at – 80 °C pending analysis. Nucleic acids were extracted from the filters and purified as described elsewhere (Winter et al. 2001, Winter et al. 2004b). In contrast to the phenol-chloroform extraction step from the original protocol, nucleic acids were extracted with 4.5 M NaCl and chloroform. This modified procedure avoids manipulation of a toxic chemical and yields fingerprints identical to those obtained by the original method (data not shown).

PCR conditions and chemicals were as described in Schäfer et al. (2001). Briefly, 1 to 4 µl of the nucleic acid extracts were quantified on an agarose gel using a DNA mass ladder (EasyLadder I; Bioline #BIO-33045). When sufficient DNA was detected it was used in 50 µl PCR reactions ($1.5 \text{ mmol L}^{-1} \text{ MgCl}_2$, 0.25 µmol L^{-1} of each primer and $2.5U \text{ Taq}$ polymerase;

Sigma; #D 5930) together with a positive and a negative control. A fragment of the 16S rRNA gene was amplified using the bacterial primer pairs 341F-GC/ 907R (Schäfer et al. 2001).

When sufficient PCR products were obtained, denaturing gradient gel electrophoresis (DGGE) was carried out as described by Schäfer et al. (2001). PCR products (500 ng) were separated into bands by electrophoresis for 18 h at 100 V on acrylamide/bis-acrylamide (8%) gels prepared using a denaturing gradient from 30% to 70% (urea and formamide). DGGE gels were photographed with a gel documentation system GelDoc EQ (Bio-Rad) after 15 min staining with a 10X SYBR Gold solution (pre-diluted in 0.2 µm filtered, autoclaved Milli-Q; Molecular Probes: # S11494). Analysis of band patterns between lanes of the same gel was performed with the Quantity One Software (Bio-Rad). Apparent bacterial richness and band intensity (measured as peak area) is considered as the number of detectable bands on the DGGE gels.

The Statgraphics Centurion XV software package (Statpoint Inc, USA) was used to test the relationships between the treatments and respective controls. After verifying that assumptions were met (Shapiro-Wilks' test), t-tests were used to determine if the effect of transplantation was significant relative to the appropriate control. Significance is given at the $p < 0.05$ level.

Results

The *in situ* physico-chemical and biological characteristics of the two sampling stations differed considerably. The freshwater (FW) station exhibited higher bacterial abundance, production, and richness, viral abundance and dissolved organic carbon concentration than in the seawater (SW) station (Table 1).

Bacterial dynamics. The initial bacterial abundance in the incubations was 5.2 and $8.3 \times 10^5 \text{ ml}^{-1}$ in the seawater and the freshwater bacterial treatments, respectively (Fig. 3A). This was

in comparison to an *in situ* abundance of 7.4×10^6 and $3.1 \times 10^6 \text{ ml}^{-1}$ for the freshwater and seawater stations, respectively (Table 1). Cell numbers continuously increased during the experiment and reached values 2 to 3 times higher at the end of the incubation. This resulted in maximum abundances that were always less than the *in situ* values by a factor of 2-3 (Table 1, Fig. 3A). The only exception was in the seawater control treatment (SB-c), where bacterial abundance declined to initial values at the endpoint.

Bacterial production increased throughout the incubation although controls and transplanted treatments showed different patterns (Fig. 3B). While a lag period occurred in the transplanted treatments, it greatly increased in the controls during the first 36 h. This trend changed thereafter and bacterial production strongly increased and was up to 3-fold higher in the transplanted treatments than in the controls. Thus, although bacterial production was low in the transplantations in the short term ($<36\text{h}$), a two-fold stimulation was observed after 48 h in SB-t and FB-t. At the end of the experiment the stimulation of bacterial production was nearly 3-fold higher in SB-t than in FB-t.

During the first 91h of the incubation, bacterial respiration was significantly lower in the transplanted treatments, relative to their respective controls (t-test, $p < 0.05$). During the second part of the incubation ($> 91\text{h}$), respiration generally decreased as a function of time in all treatments (Fig. 4A). Across the 7 day incubation period, respiration rates were statistically significantly lower in FB-t than in FB-c (t-test, $p < 0.05$). This is in contrast to the seawater bacterial transplantation (SB-t) and control (SB-c) where no significant difference (t-test, $p > 0.05$) in respiration was observed over the 7-day incubation period.

Bacterial carbon demand, which is the sum of BR and BP was relatively stable throughout the control incubations (ranging from 0.10 to $0.48 \mu\text{mol C L}^{-1} \text{ h}^{-1}$) and increased slightly at the end of the experiment (data not shown). In contrast, in the transplanted treatments, it was initially close to zero ($0.01 \mu\text{mol C L}^{-1} \text{ h}^{-1}$) but greatly increased after 2 d (up to $0.88 \mu\text{mol}$

C L⁻¹ h⁻¹) until the end of the experiment (data not shown). The stimulation of bacterial carbon demand was stronger in SB-t (up to 239%) than in FB-t (up to 112%; Fig. 4B).

The relative amount of dissolved organic carbon utilized during the incubation differed between the treatments (data not shown). There was a trend of higher consumption in the freshwater and seawater controls (22 and 13% of the initial DOC concentration, respectively) than in the corresponding transplanted treatments (9 and 3.5%, respectively).

The apparent bacterial richness (i.e. the number of bands on DGGE gels) during the incubation decreased in all treatments and this effect was detectable after 43 h or less (Fig. 5). There were significant differences between the transplanted treatments and their respective controls (t-test, $p < 0.05$). Both transplanted treatments showed a continuous decrease in apparent richness whereas it stabilized and even slightly increased towards the end in both control incubations. The largest loss in apparent richness occurred in SB-t, where a decrease of 16 bands (41%) at the endpoint of the experiment was observed, compared to 13 in FB-t and only 6 in both controls.

Transplantation effects, whether negative or positive, were always larger for SB-t than for FB-t. The relative intensity of bands differed strongly during the confinement but varied also between treatments (Fig. 6). For example, band no. 1 decreased from around 10% of the total at the start of the experiment to less than 5% towards the end of the experiments in most treatments. The relative abundance of bands 21, 23 and 27 was typically less than 5% at the start of the experiments increased over the course of the experiment, with band 21 showing large increases (up to almost 30%). Interestingly, in the FB-c the intensity of these three bands remained comparatively stable throughout the experiment.

Viral dynamics. The initial viral abundance was higher in the freshwater-diluted treatments (FB-c and SB-t) than in the seawater treatments (SB-c and FB-t) accounting for 10 x

266 10^7 and 1.5×10^7 particles ml^{-1} , respectively (Fig. 7A). This is in comparison to *in situ* values of
267 11.1 and 1.3×10^7 particles ml^{-1} for the freshwater and seawater station, respectively (Table 1).
268 Contrasting viral dynamics were observed between the treatments diluted with freshwater filtrate
269 (FB-c and SB-t), and the treatments diluted in seawater filtrate (SB-c and FB-t). The abundance
270 of freshwater viruses in FB-c and SB-t, varied more over time than the abundance of seawater
271 viruses in SB-c and FB-t, (Fig. 7A) and strongly decreased during the first day of incubation (by
272 33% and 42% in FB-c and SB-t, respectively). At the end of the experiment, the viral abundance
273 in FB-c and SB-t was slightly lower than at T_0 (15%). In FB-t viral abundance at the endpoint
274 was similar to T_0 , while in SB-c it was 60 % higher than the initial values.

275 Although viral production exhibited similar patterns during the control incubations, it was
276 43% to 235% higher in FB-c than in SB-c. Total viral production over the entire experiment was
277 lower than decay in FB-c, FB-t and SB-t and viral production significantly exceeded the decay
278 rate only in SB-c. By comparing the viral dynamics in FB-t and SB-t, relative to the seawater
279 and freshwater controls (i.e. FB-t compared to SB-c and SB-t compared to FB-c), respectively, it
280 is possible to evaluate the response of freshwater and marine viruses to the presence of a non-
281 indigenous bacterial community (Fig. 7B). Despite initially high viral production values, viral
282 production was repressed by up to 190% in the FB-t incubation, relative to the seawater control
283 (SB-c). In contrast, the transplantation of a seawater bacterial community (SB-t) into freshwater
284 viral community resulted in a strong stimulation of viral production (up to 840% after 91h).

285 The percentage of bacterial cells lysed per day can be estimated from the net increases in
286 viral abundance in the incubations and assuming a burst size of 50 viruses, which represents high
287 values from the North Sea (Winter et al. 2004a). Bacterial mortality due to viruses was higher in
288 FB-c than in SB-c (23% and 12% cells lysed per day, respectively) and while transplantation of
289 seawater bacteria increased virus-induced mortality to $52\% \text{ d}^{-1}$, the transplantation of freshwater
290 bacteria decreased mortality to $10\% \text{ d}^{-1}$.

291 Most viruses (50-83%) were from the low fluorescence group (V1). However, the relative
292 contribution of V1 to the total abundance differed between treatments (Fig. 7C). Over the course
293 of the experiment, the contribution of the V1 group varied little in the SB-c and FB-t treatments.
294 This is in contrast to the SB-t incubation, where the proportion of the V1 group was between that
295 of FB-c (virus source) and SB-c (bacteria source) and exhibited an increasing trend towards the
296 end of the incubation.

DISCUSSION

Metabolic and structural response of bacterial communities to transplantation.

This study investigates how the metabolism and diversity of estuarine and riverine bacteria responds to transplantation and to exposure to non-indigenous virus communities. The results show that transplantation increased both bacterial production and carbon demand in seawater as well as freshwater bacterial communities. This response is concomitant with a decrease in apparent bacterial richness. It suggests that bacteria were able to grow under allochthonous sources of dissolved organic matter supporting the conclusions of a previous study carried out in a Danish estuary (Rochelle-Newall et al. 2004). The data also show that transplantation stimulated total virus production and virus-induced mortality in SB-t but repressed it in FB-t.

It is well established that incubations can significantly affect bacterial community composition as well as activity parameters (Massana et al. 2001, Gattuso et al. 2002, Winter et al. 2004b). However, incubations are often the only possible approach to investigate ecological and biogeochemical issues. In the present study, the decrease in bacterial richness in the control incubations reflects both a confinement effect and the impact of the initial dilution. The decline of richness was similar in the two controls, where 6 bands were lost by the end of the experiment. In contrast, band loss was more than twice as high in the transplanted communities (13 and 16 bands for the FB-t and SB-t communities, respectively). This implies that the negative effects of transplantation on bacterial richness were higher than those of confinement alone. The lower apparent richness is probably related to the higher dominance of a few favored, fast growing phylotypes in the transplanted treatments relative to the controls. Furthermore, the loss of some of the phylotypes may well have helped to enhance the dominance of faster growing ones, such as can be observed in the SB-t incubation with loss of band 15 and the increase in band 23 intensity throughout the experiment.

These opportunistic species may have taken advantage of the nutrient amendment resulting from the dilution thus out-competing other members of the assemblage (Massana et al. 2001, Gattuso et al. 2002, Winter et al. 2004b). Also, exposure of seawater bacteria to freshwater dissolved organic matter (FW-DOM) could have activated specific ectoenzymes or have favored certain members of the community capable of expressing them for hydrolyzing allochthonous DOM (Kirchman et al. 2004). This could explain the decrease in richness and the changes in community composition (Pinhassi et al. 1999).

The increase in bacterial production, respiration and carbon demand in the transplantations indicates that seawater bacteria could readily oxidize freshwater dissolved organic carbon, despite the fact that it has been reported as relatively less labile than marine DOC (del Giorgio & Davis 2003). This argument, combined with the relatively long residence time in the Scheldt estuary (1 to 3 months according to Wollast (1988), suggests that it is unlikely that there was an export of a labile FW DOC fraction that had not been completely taken up by the riverine bacteria as observed in other areas (Kirchman et al. 2004).

The higher metabolic activity and larger decrease in bacterial richness observed in SB-t compared to FB-t may also result from the different virus-induced bacterial mortality. Indeed, the virally-induced mortality was almost 5 times higher in the SB-t than in the FB-t. Moreover, although the change in salinity experienced by the FB-t treatment could have had a negative effect on the bacterial communities there was no salinity effect for the seawater bacteria transplanted into freshwater DOM as salinity was adjusted.

Viral dynamics. The experimental setup (dilution of bacterial communities with $< 0.2 \mu\text{m}$ filtered water) artificially increased the initial virus to bacteria ratio. It was 5 to 9 times higher in the controls than in the corresponding inoculum. It increased by a factor of 9 in SB-t but was 5

times lower in FB-t as compared to FB-c. The FB-t bacteria therefore experienced reduced viral pressures relative to FB-c whereas SB-t bacteria experienced a stronger viral pressure than SB-c.

It has been shown that viral communities from different parts of an estuary exhibit pronounced differences in the genotypic composition (Wommack et al. 1999). Thus, the viral communities in the freshwater and seawater parts of the estuary should differ and this is potentially supported by the difference in the proportions of viral groups, as measured by fluorescence, between the two environments. There were differences between the FB-c and SB-c treatments potentially suggesting the presence of different viral communities in seawater and freshwater. These differences in relative proportion of low fluorescence viruses were also evident in the FB-t and SB-t incubations, reflecting the seawater and freshwater viral sources, respectively. The transplantation of seawater bacteria into a freshwater viral community resulted in large changes in the relative proportion of the low fluorescence virus group which may suggest that the seawater bacterial community produced viruses which differed from viruses produced by the indigenous freshwater bacterial community. However, this remains to be clarified in a more rigorous manner. Nevertheless, this hypothesis is also supported by the fact that the SB-t transplantation resulted in an elevated viral production rate (Fig. 7B). Thus, it is possible that the transplanted seawater bacteria were infected by viruses of the freshwater community.

Similar observations have been made by Sano et al. (2004). Interestingly, this pattern did not seem to hold for the seawater viruses in the FB-t incubation. Although viral production was high in the first hours, viral production was much lower in the transplanted incubation than in the control. This is suggestive of an increased survival of transplanted bacteria relative to that of the control and points towards the hypothesis that the freshwater bacteria did not act as hosts for the seawater viruses. Although the mechanisms for the lower viral production of the freshwater bacteria are not clear, the salinity effects of transplanting freshwater bacteria into virus-

containing seawater may play a non-negligible role in altering the virus-bacterial encounter and infection rates.

It should be noted that we compared treatments with similar initial abundances (FB-t with SB-c and SB-t with FB-c). Within these comparisons, total contact rate between viruses was 60% higher in the treatment with freshwater bacteria since initial bacterial abundance was higher in freshwater and the same % dilution was used for freshwater and seawater bacteria in the experiments. This could result in higher viral production rates by the freshwater bacteria (Murray & Jackson 1992). However, we found the opposite trend, i.e. transplantation of seawater bacteria into freshwater resulted in higher net viral production than for the freshwater bacteria, whereas transplantation of freshwater bacteria into seawater resulted in lower net viral production than the seawater control (Fig. 7B).

Viral production could be detected in the transplanted freshwater and the transplanted seawater bacterial community (Fig. 7A). Several non-mutually exclusive mechanisms could explain the production of viruses in transplanted bacterial communities. Firstly, the host range could be broader than previously assumed, allowing viruses to infect hosts from different environments. For example, Jensen et al. (1998) have argued that the concept derived from isolated virus-hosts systems that viruses do not trespass the genus barrier (Ackerman & Dubow 1987) is an isolation artifact. This is also supported by Chiura (1997) who has shown that marine viruses can infect *Escherichia coli*. Secondly, transplantation into a different environment (e.g. with a large change in salinity) could cause induction of lysogenic cells (Jiang & Paul 1996). The enhanced growth rates (Fig. 3 and 4) could have acted as inducing agent (Weinbauer 2004) and caused the virus production observed. Finally, it is possible that cosmopolitan bacterial phylotypes that can grow and produce viruses in freshwater and marine conditions exist. Indeed, several identical bands were found in both environments and all treatments. Although there is no definitive evidence to support the hypothesis that freshwater viruses can infect seawater bacteria

396 and vice versa, it is clear that transplanted bacteria were able to produce viruses in the new
397 environment, at least in the freshwater virus incubations. Moreover, as we found a strong
398 increase in virus production, it is clear that these viruses came from either the transplanted
399 community (by induction) or by new infection from the original viral community. Obviously,
400 our estimates of viral production and decay can only be considered as net changes and so must
401 be viewed as conservative estimates. Nevertheless, it is clear that these estimates were different
402 between the different incubations (e.g. strong stimulation in SB-t and strong repression in FB-t)
403 and so it is probable that there were real differences in viral production and decay between the
404 different treatments. To conclude, although the freshwater viral community seems to be able to
405 infect seawater hosts, the converse did not appear to occur.

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Table 1. Physico-chemical and biological characteristics of surface water sampled at the freshwater (FW) and seawater (SW) ends of the Scheldt estuary. DOC: dissolved organic carbon \pm standard deviation. BP: bacterial production \pm standard deviation. DGGE: Denaturing Gradient Gel Electrophoresis.

	Stations	
	FW	SW
Temperature (°C)	11.04	7.73
Salinity	1.20	30.4
DOC ($\mu\text{mol l}^{-1}$)	427 ± 1.9	195.0 ± 1.6
Virus ($\times 10^7 \text{ ml}^{-1}$)	11.1	1.3
Bacteria ($\times 10^6 \text{ ml}^{-1}$)	7.4	3.1
BP ($\mu\text{mol C l}^{-1} \text{ h}^{-1}$)	0.255 ± 0.006	0.070 ± 0.002
Bacterial richness (# of DGGE bands)	40	37

Figure legends

Figure 1. Map of the estuary and sample sites

Figure 2. Experimental set-up. Seawater or Freshwater (60 L) was filtered sequentially through GF/F (Whatman) filters and 0.2 μ m Durapore (Millipore) cartridge filters. 18 L of each filtrate was dispatched into four 20 L polycarbonate carboys. 1 L of unfiltered freshwater or seawater was added to each incubation to provide the natural bacterial inoculum. Two transplants and two controls were used: SB-c: seawater + seawater bacteria; SB-t: freshwater + seawater bacteria; FB-c: freshwater + freshwater bacteria; FB-t: seawater + freshwater bacteria.

Figure 3. Bacterial abundance (BA; panel A) and production (BP; panel B) in the different treatments: SB-c: seawater + seawater bacteria; SB-t: seawater bacteria transplanted in freshwater; FB-c: freshwater + freshwater bacteria; FB-t: freshwater bacteria transplanted in seawater. In panel B values represent mean \pm standard deviation

Figure 4. Bacterial respiration (BR) in the four treatments during the different phases of the experiment (panel A). Transplantation effect on bacterial carbon demand (BCD) across the incubation period in transplanted treatments (SB-t and FB-t) relative to their corresponding controls (panel B).

Figure 5. DGGE gels from the incubations. Each lane is labeled with the incubation and the time point. The bands discussed in the text are also noted. Upper image: time points T0 to T43, lower image T91 to T157. SB-c: seawater + seawater bacteria; SB-t: seawater bacteria transplanted in freshwater; FB-c: freshwater + freshwater bacteria; FB-t: freshwater bacteria transplanted in seawater. SD: standard.

Figure 6. Bacterial richness in the four treatments.

Figure 7. Viral abundance (VA; panel A), relative virus production (panel B) and relative % of V1 (low fluorescence group) in the different treatments across the experiment. Relative virus production is expressed as a % of the respective viral control (SB-c for the FB-t and FB-c for SB-t). The relative percentage of V1 is expressed as the % of the total abundance. SB-c:

598 seawater + seawater bacteria; SB-t: transplanted seawater bacteria; FB-c: freshwater +
599 freshwater bacteria; FB-t: transplanted freshwater bacteria.

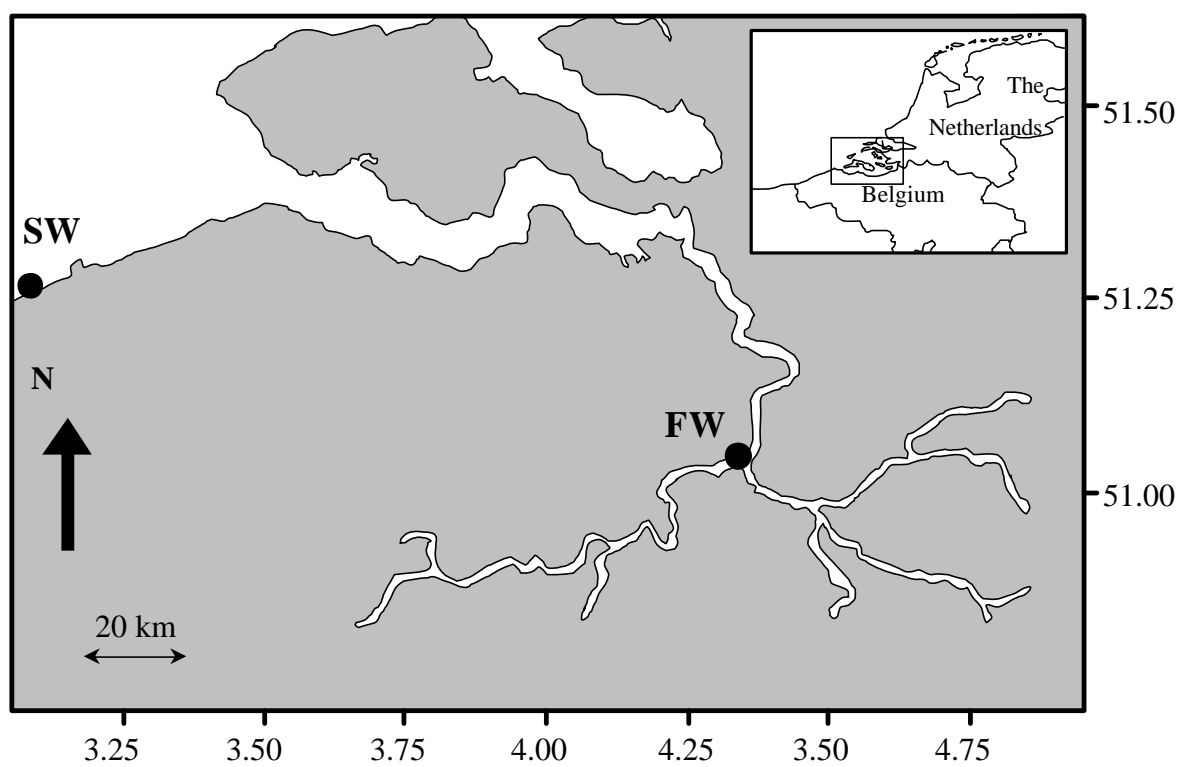


Fig. 1

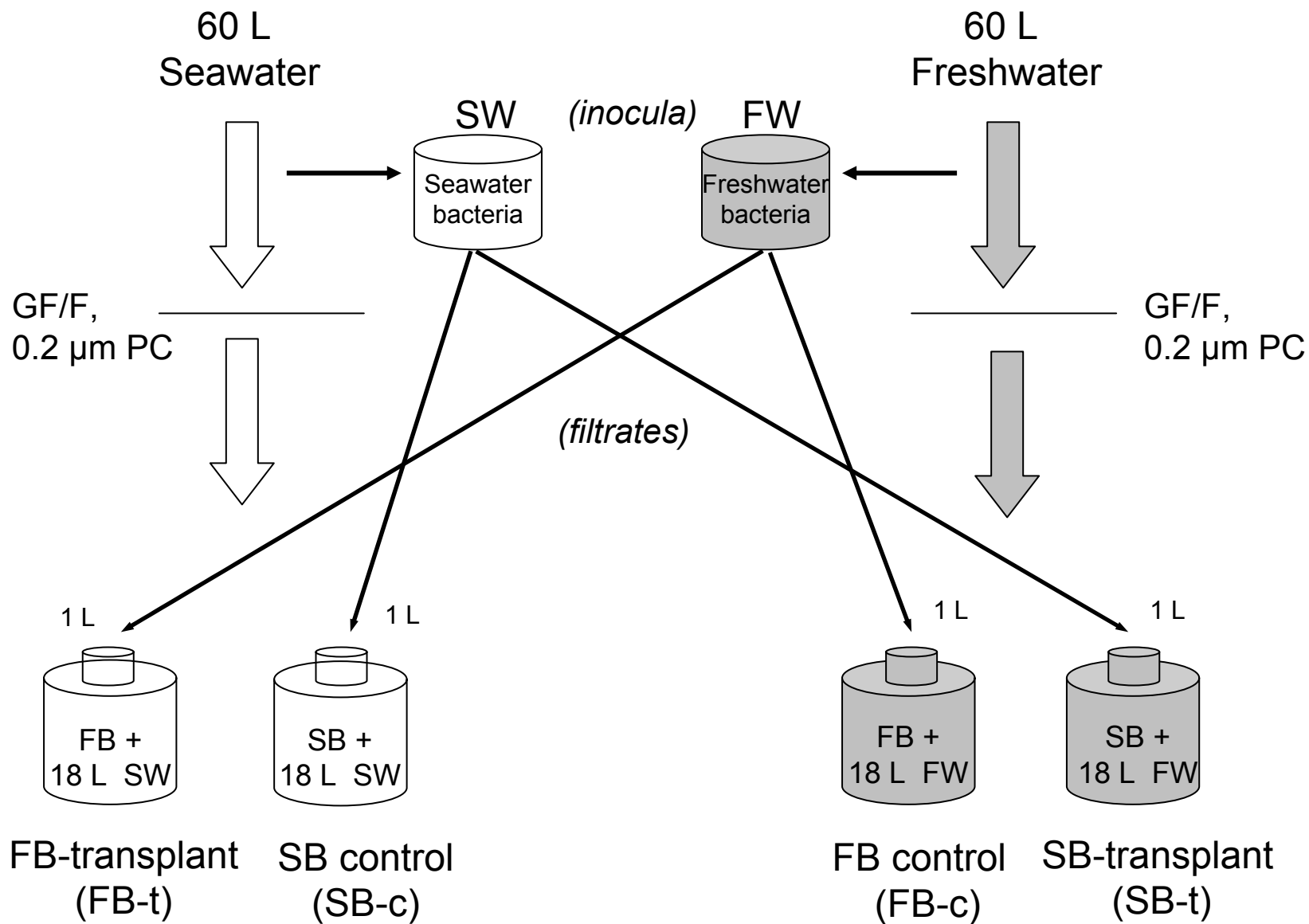


Fig. 2

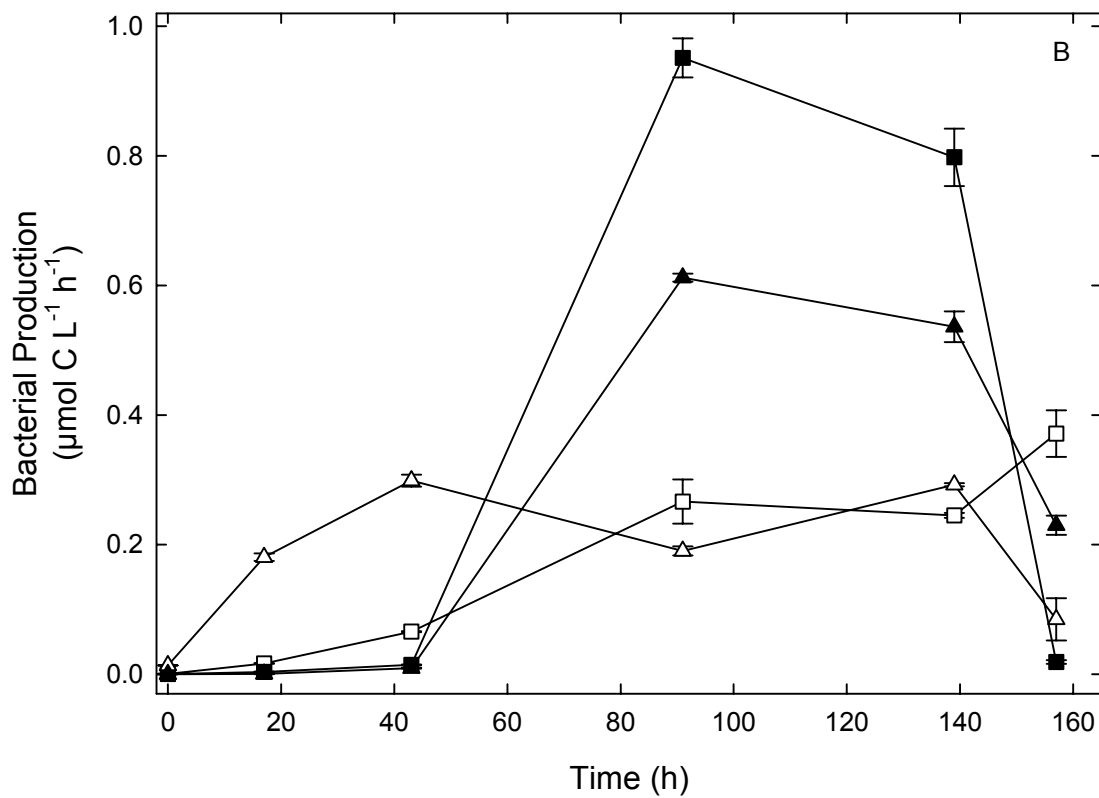
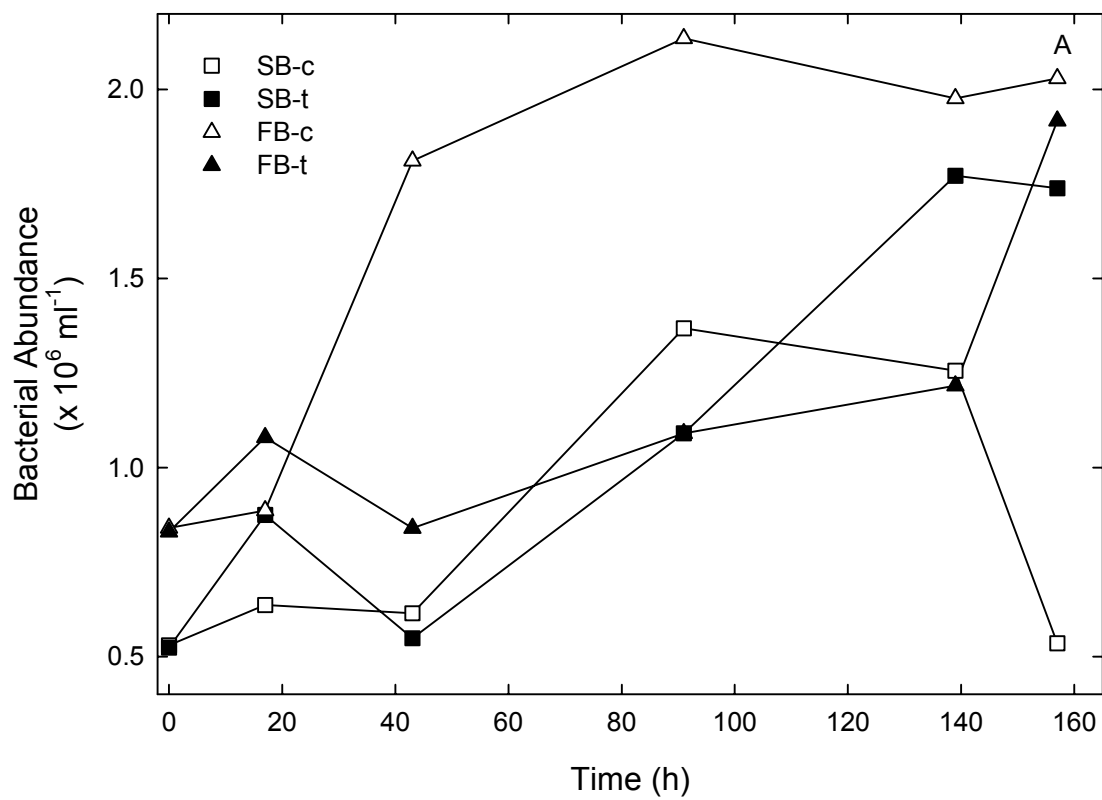


Fig. 3

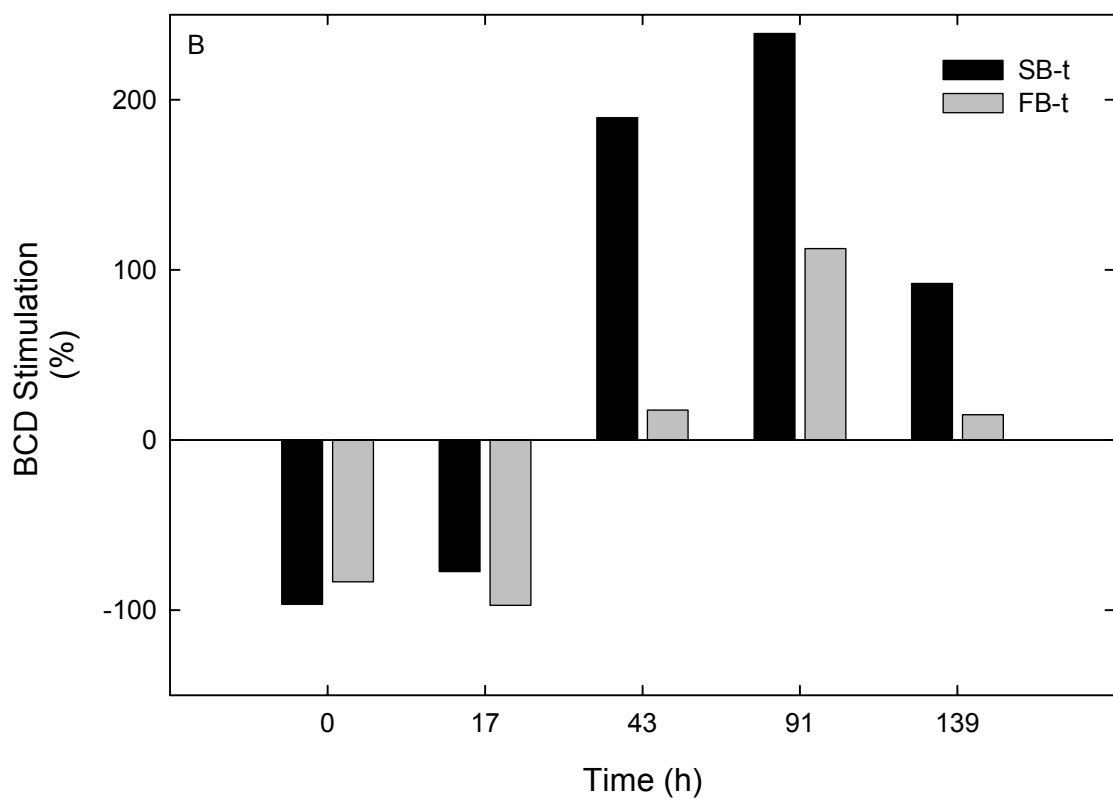
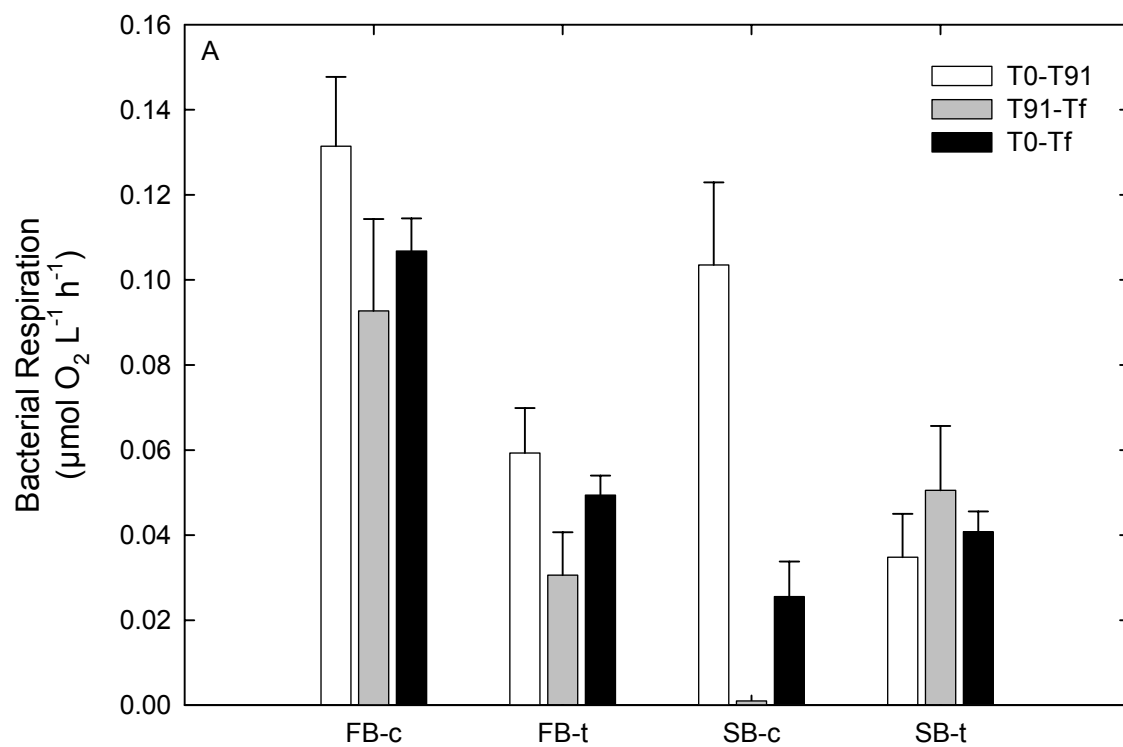


Fig. 4

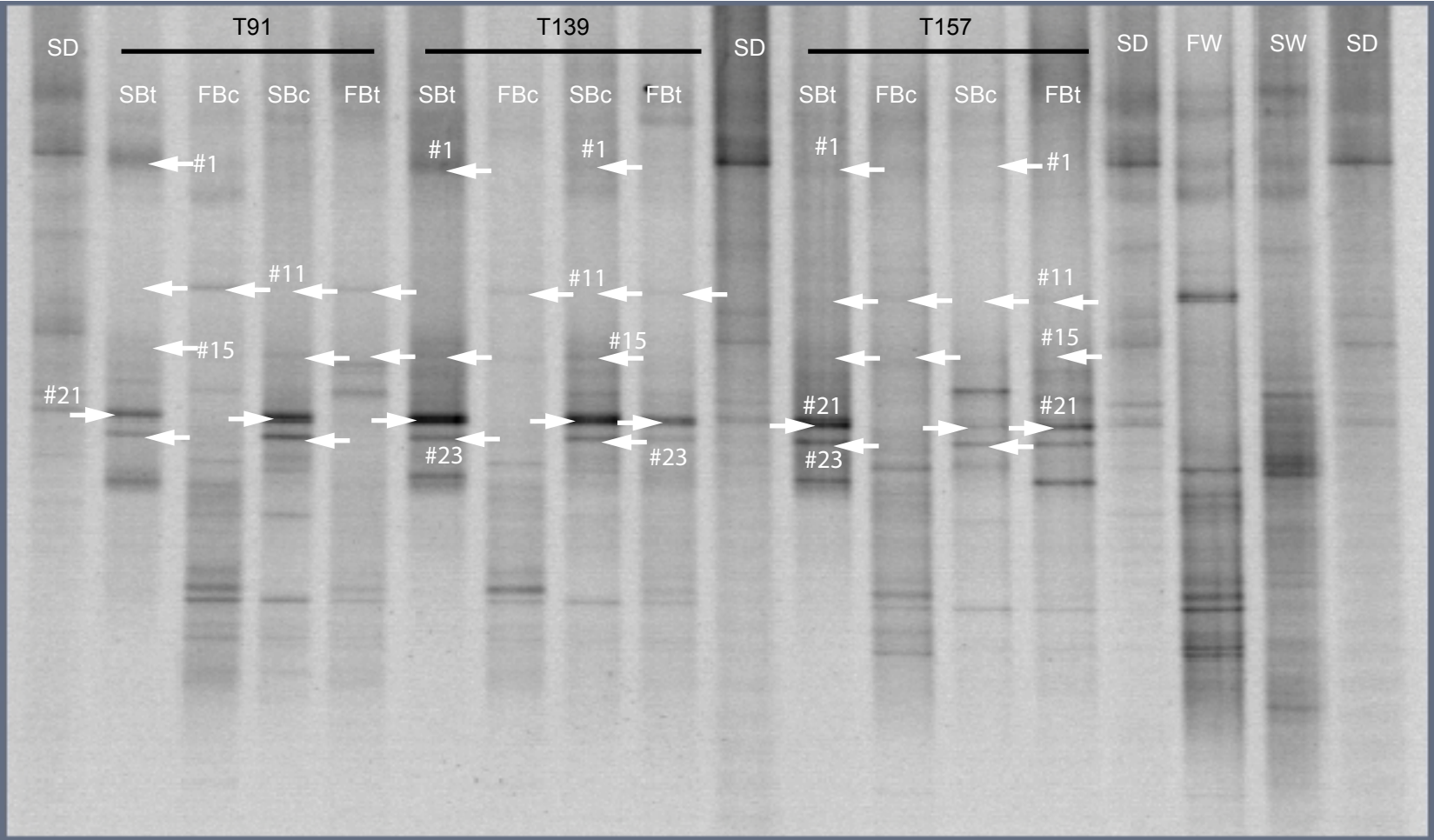
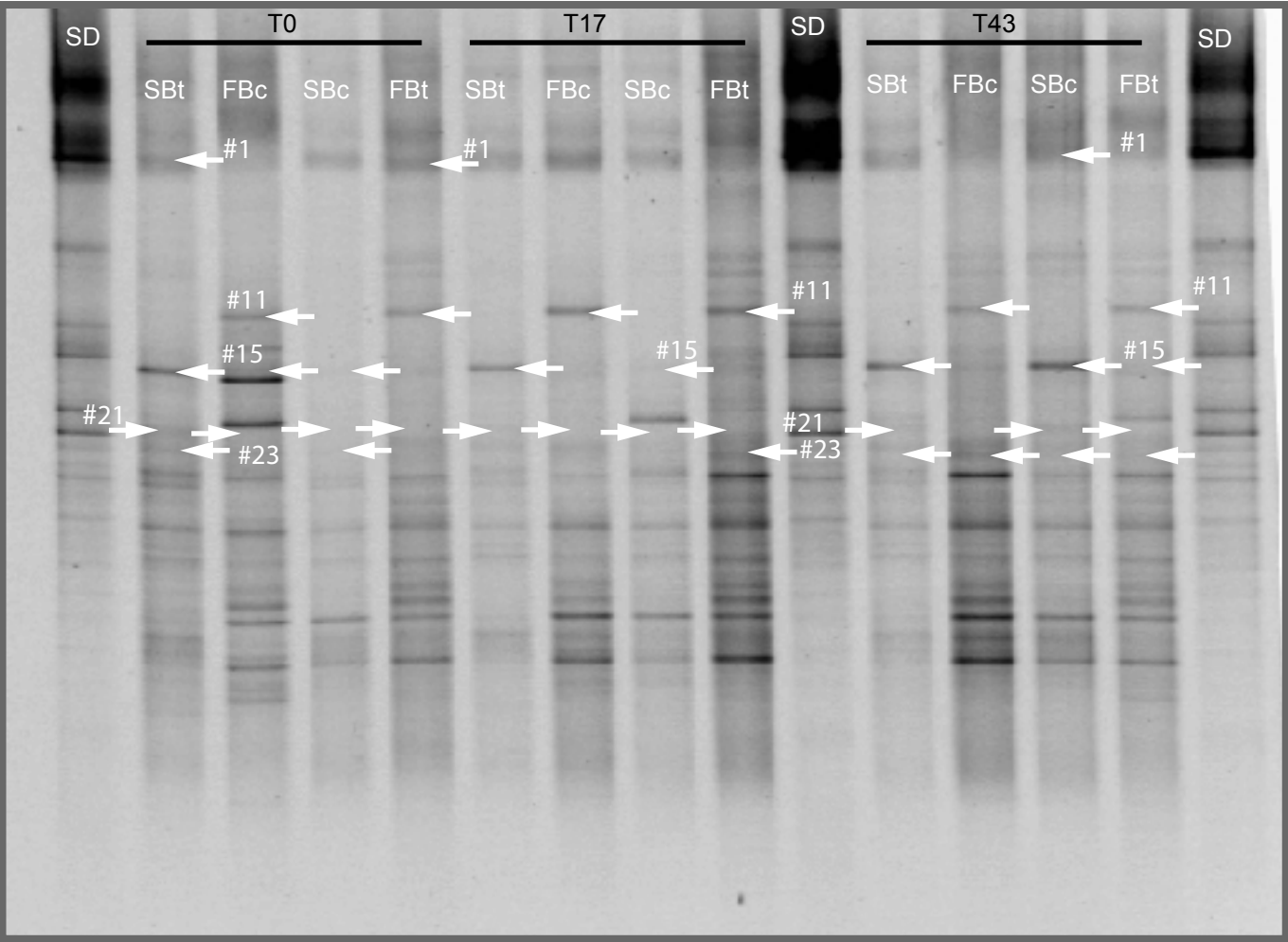


Fig. 5

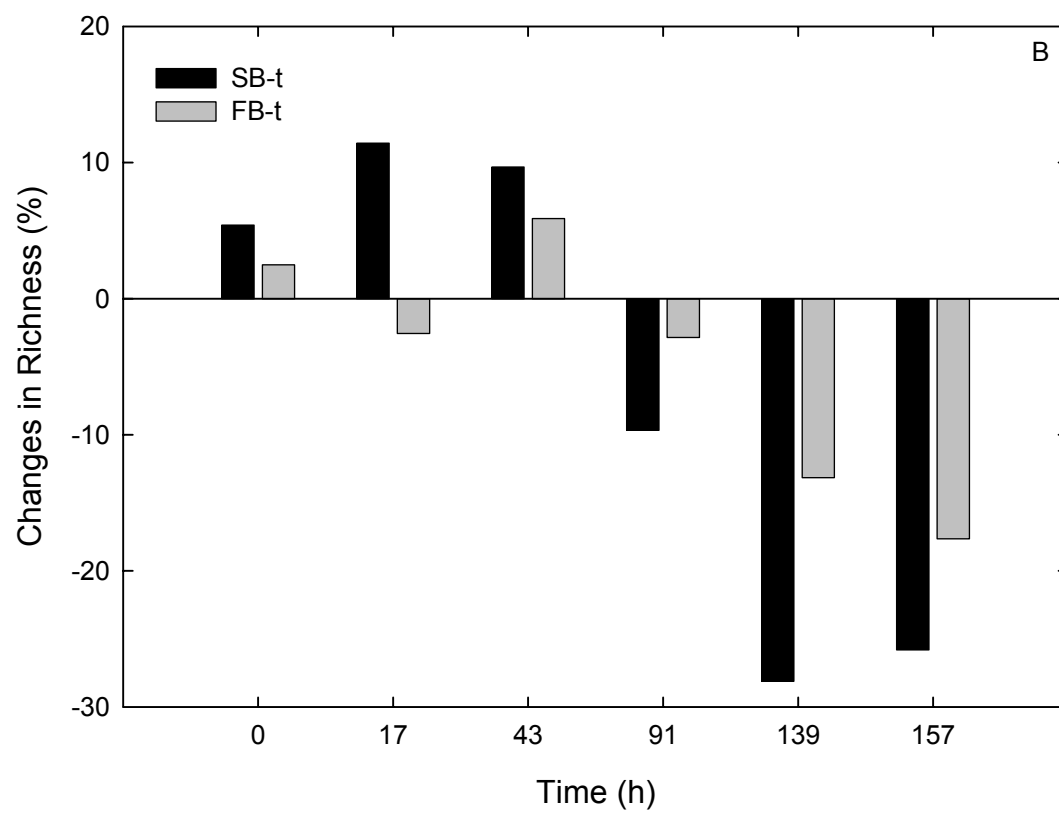
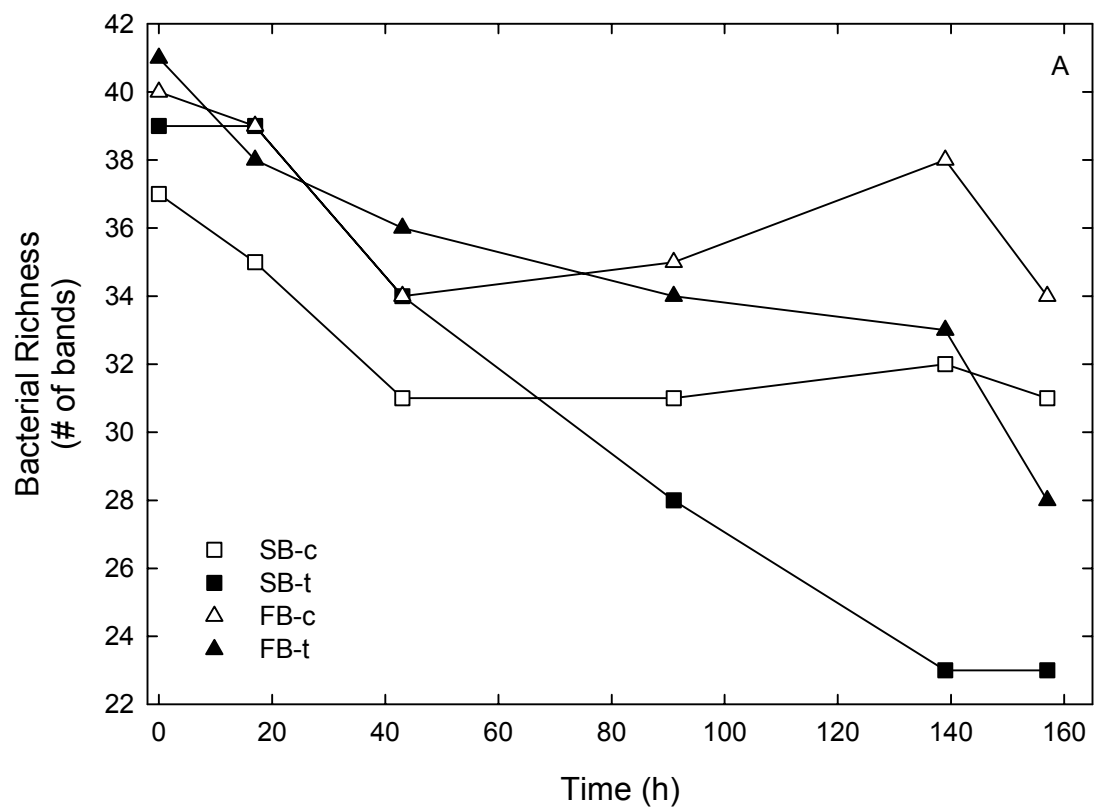


Fig. 6

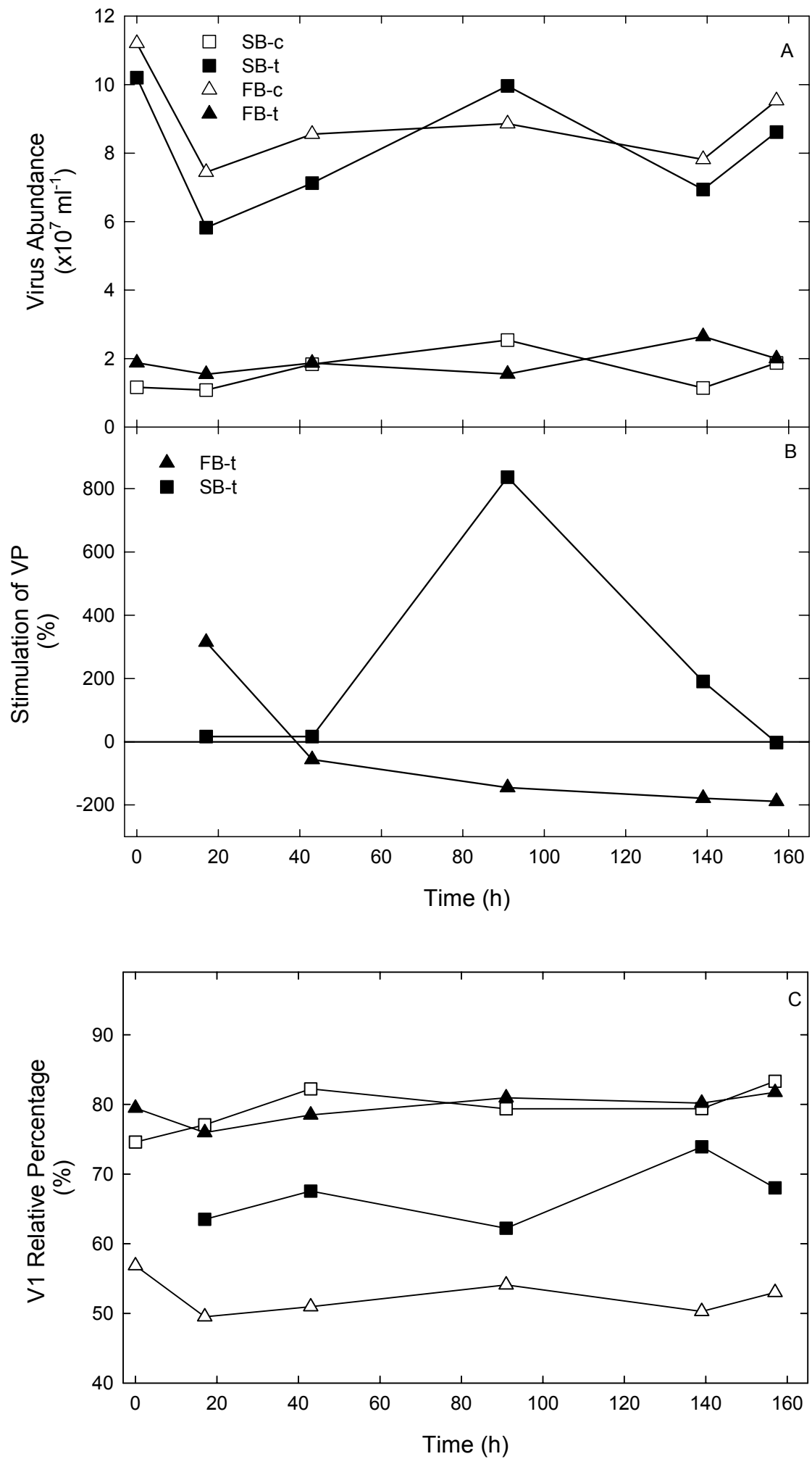


Fig. 7