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Earthworm effects on plant growth do not necessarily decrease with soil fertility

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Abstract Earthworms are known to generally increase plant growth. However, because plant-earthworm interactions are potentially mediated by soil characteristics the response of plants to earthworms should depend on the soil type. In a greenhouse microcosm experiment, the responsiveness of plants (*Veronica persica*, *Trifolium dubium* and *Poa annua*) to two earthworm species (in combination or not) belonging to different functional groups (*Aporrectodea caliginosa* an endogeic species, *Lumbricus terrestris* an anecic species) was measured in term of biomass accumulation. This responsiveness was compared in two soils (nutrient rich and nutrient poor) and two mineral fertilization treatments (with and without). The main significant effects on plant growth

were due to the anecic earthworm species. *L. terrestris* increased the shoot biomass and the total biomass of *T. dubium* only in the rich soil. It increased also the total biomass of *P. annua* without mineral fertilization but had the opposite effect with fertilization. Mineral fertilization, in the presence of *L. terrestris*, also reduced the total biomass of *V. persica*. *L. terrestris* did not only affect plant growth. In *P. annua* and *V. persica* *A. caliginosa* and *L. terrestris* also affected the shoot/root ratio and this effect depended on soil type. Finally, few significant interactions were found between the anecic and the endogeic earthworms and these interactions did not depend on the soil type. A general idea would be that earthworms mostly increase plant growth through the enhancement of mineralization and that earthworm effects should decrease in nutrient-rich soils or with mineral fertilization. However, our results show that this view does not hold and that other mechanisms are influential.

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Introduction

Soil organisms are known to affect plant growth by enhancing mineralisation of soil organic matter and modifying physical and chemical properties of soil

(Bardgett et al. 2005; Lavelle and Spain 2001). Within soil organisms, earthworms are in term of biomass and activity among the most important detritivores in terrestrial ecosystems (Edwards 2004). They are also known to affect plant growth, generally positively, via five main mechanisms (Brown et al. 2004; Scheu 2003): (1) an increased mineralization of soil organic matter (2) the production of plant growth substances via the stimulation of microbial activity; (3) the control of pests and parasites; (4) the stimulation of symbionts and (5) modifications of soil porosity and aggregation, which induces changes in water and oxygen availability to plant roots. Although these mechanisms are well identified it is difficult to determine their relative influence (Blouin et al. 2006) either in precise cases or in broad classes of cases as defined by plant functional type, geographic area or soil type.

Brown et al. (2004) has remarked that the response of plants to earthworms should depend on the type of soil and especially its texture and its richness in mineral nutrients and organic matter. Indeed, mechanisms through which earthworms influence plant growth might be either down or up-regulated by soil characteristics. For example, if the positive effect of an earthworm species on plant growth is mainly due to an increase in mineralization, the species might no longer increase plant growth in a soil where nutrients are not limiting. However, few studies (Doube et al. 1997; Wurst and Jones 2003) have tested the effect of earthworms on plant in different soils in the same laboratory experiment. Doube et al. (1997) have shown that *Aporrectodea trapezoides* increased the growth of wheat in a sandy soil but not in a clayey one. They also showed that the growth and the grain yield of barley were increased by *Aporrectodea trapezoides* and *Aporrectodea rosea* (both are endogeic earthworms) in the sandy soil but reduced in the clayey. On the contrary, Wurst and Jones (2003) have shown that *Aporrectodea caliginosa* increased the root biomass of *Cardamine hirsute* in two different soils. Up to our knowledge, no laboratory experiment has so far compared the effect of two earthworm species belonging to different functional groups on the same plant species, in soils differing by their texture and nutrient content. Our experiment aims at meeting this need.

Besides, plant species of different functional groups should respond differently to earthworms

(Brown et al. 2004) because they are not limited by the same resources and do not have the same resource allocation strategies (Laossi et al. 2009). Legumes, for example, are thought to be less responsive to earthworms than grasses since they are not limited by nitrogen (Brown et al. 2004; Wurst et al. 2005). Finally, plant responses may also depend on earthworm functional group since earthworm belonging to different functional groups differ in their behaviors (Lavelle and Spain 2001). For example, endogeic earthworms keep moving inside the soil to feed on soil organic matter while anecic feed on plant litter at the soil surface and tend to stay in the same burrow (Lavelle et al. 1998). Anecic earthworms fragment plant litter and incorporate it into the soil where it can subsequently be ingested by endogeic earthworms. Such an interaction could increase further mineralization and plant growth (Brown et al. 2000; Jégou et al. 1998).

We tested how the effects of earthworms belonging to different functional groups on plant growth covary with plant functional group and soil type. Hence, we investigated in a microcosm greenhouse experiment the responsiveness of three annual plant species of different functional groups (*Poa annua*, a grass; *Trifolium dubium*, a legume; *Veronica persica*, a forb) to an endogeic (*Aporrectodea caliginosa*) and anecic (*Lumbricus terrestris*) earthworm species as well as to the combination of the two species. The responsiveness of plant was measured in term of biomass accumulation and was compared in two soils (a clayey nutrient rich soil with higher organic matter content and a sandy nutrient poor soil with lower organic matter content) and two mineral fertilization treatments (with and without). Mineral fertilization can be considered either to mimic richer soils or agricultural practices.

We hypothesized that earthworms affected plant growth mainly through an enhancement of mineralization, which is the more often cited mechanism (Kreuzer et al. 2004; Patsch et al. 2006; Wurst et al. 2003). Therefore (see above), both earthworm species should affect (1) plant growth and (2) plant resource allocation only in the soil that is poor in organic matter and mineral nutrient. Similarly, (3) significant interactive effects of the two earthworm species on plant growth and resource allocation should only be found in the poor soil. Finally, according to the same assumption that earthworm mostly influence plants by increasing the availability of mineral nutrients, (4) the

impact of *A. caliginosa* and *L. terrestris* on plants should decrease with mineral fertilization.

Materials and methods

Experiment set up

The experiment was set up in microcosms consisting of PVC pots (inner diameter 14 cm, height 12.5 cm) that were closed at the bottom with 1 mm plastic mesh to prevent earthworms from escaping. A total of 320 microcosms were filled with 950 g (± 20 g) of sieved (2 mm) dry soil in a greenhouse. Before starting our experiment, the microcosms were watered regularly for 2 weeks and germinating weeds from the seedbank were removed. Eight grams of dried litter (72 h at 60°C) of grass leaves were placed at the soil surface and 1 g was mixed with the first cm of soil, prior to the addition of earthworms and seeds. This constituted the essential food resource for the anecic earthworm species.

Soils

We used two different soils: A sandy cambisol (called hereafter in the text the “nutrient poor soil”) supporting a meadow (OM=2.55%; C/N ratio=12.4; total carbon content=1.47%; Ntotal=0.12%; pH=5.22) collected at the ecology station of the Ecole Normale Supérieure at Foljuif (France) and a clayey leptosol (OM=9.81%; C/N ratio=12.2; total carbon content=5.67; Ntotal=0.465; pH=7.45) collected at the ecology station of Brunoy (France) (called hereafter the “nutrient rich soil”).

Earthworms

We used an anecic earthworm, *Lumbricus terrestris* (L.) -LT-, and an endogeic earthworm, *Aporrectodea caliginosa* (Savigny) -AC-. LT were purchased in a store and AC were collected in the park of the IRD centre in Bondy (France). Our experiment consisted in four treatments: AC, LT, AC + LT and a control without any earthworm species (C). One adult of LT (4.2 ± 0.5 g) and three adults of AC (2.4 ± 0.4 g) were introduced in each treatment including these species. This represents respectively 273 g m^{-2} and 156 g m^{-2} , which is comparable to the biomasses found in temperate grassland ecosystems (Edwards and Bohlen

1996). In the treatment with both earthworm species (AC + LT) we have maintained for each earthworm species the biomass used in AC and LT treatment. This was done to maintain the same activity level of each earthworm species, which was the only way to allow testing for a possible interactive effect of the two earthworm species on plant growth. 96% of the earthworms were recovered at the end of the experiment (471 *A. caliginosa* and 143 *L. terrestris* individuals among the 480 and 160 that were originally introduced respectively).

Plants

One week after introducing earthworms, 15 seeds of *Veronica persica*, *Trifolium dubium* and *Poa annua* were sown in monocultures. Three weeks later, a single plant per microcosm was kept (the other seedlings were removed and cut down in the original microcosm). Microcosms were weeded every week during the experiment. Microcosms were watered during 7 weeks with 6.5 ml every day and from 8th week to the end (week 16) with 13 ml every day. This allowed us to maintain the soil near its field capacity (this was checked through regular weighing of some pots). Microcosm position within the greenhouse was randomized every 2 weeks.

Fertilizer

For each combination of treatments (soil type \times LT \times AC \times plant species), two fertilization treatments were implemented: without or with mineral fertilization. This treatment consists in an application of fertilizer containing N, P, K, S and Mg. 0.6 g of fertilizer was placed at the soil surface at the beginning of the experiment, 0.6 g 3 weeks after sowing and 0.6 g on week 6, when the first flowers were produced. From week 6 to week 12, 1 g of fertilizer was added every 2 weeks before watering. A total of 5.8 g of fertilizer was then added per pot. This corresponds to 48 kg of N and K; 32 kg of P; 6.7 kg of S and 97 kg of Mg per hectare. Five replicates of each treatment combination, i.e. fertilization \times soil type \times LT \times AC \times plant species, were implemented.

Sampling

Plants were harvested on week 16. Shoot (leaves and stems) were cut at the soil surface and roots were separated from the soil by washing on a 600 μm mesh.

Root and shoot biomasses were dried at 60°C for 72 h. Dried shoot and root biomasses were weighted. Because of differences in the timing of seed maturation between plant species, seeds were not harvested.

Statistical analyses

Data were analysed with ANOVAs using SAS GLM procedure (Sum of squares type III, SS3) (SAS 1990). A full model was first used to test all factors (“AC”, “LT”, “plant species”, “fertilizer” and “soil”) and all interactions between them (Table 1). When significant interactions between plant species and other factors (AC, LT, soil and fertilization) were detected, data were reanalysed separately for each plant species (Table 2) to describe in a more detailed way the effects of these treatments on each plant species. This allowed for example determining which plant species responded to which earthworm species and in which

conditions (soil type and fertilization). Effects of treatments and interactions between treatments were tested on shoot biomass, root biomass, total biomass, and shoot/root ratio.

The residuals of each model were analysed to test for normality and homogeneity of variances. To determine the direction of significant effects, we used multiple comparison tests based of least square means taking into account Bonferroni’s correction (LSmeans, LSmeans SAS statement). All tests were achieved with a significance level $\alpha=0.05$.

Results

In the present experiment, the main effects are due to the anecic earthworm species, *L. terrestris*. No significant effect of *A. caliginosa* was found. The full statistical model (Table 1) showed that *L. terrestris*

Table 1 ANOVA table of *F*-values for the effects of earthworms (AC and LT), soils, fertilizer and plant species on root, shoot and total biomass and shoot/root ratio (Total df=209)

	df	Root biomass	Shoot biomass	Total biomass	Shoot/root
		F	F	F	F
AC	1	0.63	2.72	1.83	2.90
LT	1	0.30	0.03	0.07	6.63*
Soil	1	4.30*	18.14***	12.17***	44.31***
Fertilizer	1	42.92***	92.65***	55.68***	139.48***
Plant species	2	199.22***	536.48***	609.26***	6.07**
AC*LT	1	0.01	1.20	1.08	0.00
AC*soil	1	3.14	0.11	0.03	0.03
AC*fertilizer	1	1.48	0.00	0.06	0.10
AC*plant species	2	2.53	0.10	0.05	1.95
LT*soil	1	0.11	4.02*	3.47*	0.00
LT*fertilizer	1	3.19	10.57***	11.46***	4.09*
LT*plant species	2	0.90	4.56**	4.45**	0.01
Soil*plant species	2	8.65***	17.34***	20.46***	1.10
Soil*fertilizer	1	10.70***	4.00*	1.26	42.48***
Fertilizer*plant species	2	31.48***	27.74***	14.01***	0.94
AC*LT*soil	1	0.11	0.08	0.11	0.57
AC*LT*plant species	2	0.59	0.32	0.25	6.07***
AC*LT*fertilizer	1	0.10	0.99	1.00	0.43
LT*soil*fertilizer	1	0.07	0.54	0.39	0.02
AC*soil*fertilizer	1	0.40	1.00	0.62	0.98
plant species*soil*fertilizer	2	1.07	0.78	1.07	1.26
AC*LT*soil*fertilizer	1	0.15	0.56	0.61	0.42
AC*LT*plant species*fertilizer	2	3.12	68.18***	72.46***	5.08*
r ²		0.79	0.90	0.89	0.62

* $p < 0.05$; ** $p < 0.01$;

*** $p < 0.001$

Table 2 ANOVA table of *F*-values for effects of earthworms (AC and LT), soil and fertilizer on root, shoot and total biomass and shoot-to-root ratio of *Trifolium dubium* (total df=78), *Poa annua* (total df=74) and *Veronica persica* (total df=55)

df	T. dubium						P. annua						V. persica											
	Root biomass		Shoot biomass		Total biomass		Shoot/ root ratio		Root biomass		Shoot biomass		Total biomass		Shoot/ root ratio		Root biomass		Shoot biomass		Total biomass		Shoot/ root ratio	
	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	F	
LT	.	9.10**	10.95**	1.34	.	0.4	4.66*	0.31	0.4	0.38	0.42	4.91*	0.31	0.4	0.31	4.66*	0.31	0.4	0.38	0.42	4.91*	0.31	0.4	0.31
Soil	35.71***	117.69***	135.03***	11.56**	5.93*	2.43	24.05***	1.04	5.93*	13.66**	14.74***	1.81	1.04	1.04	1.04	24.05***	1.04	1.53	6.22**	6.65**	6.72	1.53	1.53	1.53
Fertilizer	29.40***	7.99**	4.79*	29.63***	99.80***	87.66***	153.33***	48.25***	99.80***	1.27	0.40	19.94***	4.69	48.25***	48.25***	153.33***	4.69	15.41***	0.03	0.20	0.03	0.07	15.41***	15.41***
AC*LT	.	.	.	3.18**	.	.	0.31	0.31	.	.	.	0.31	5.68**
LT*soil	.	9.54**	10.11**	1.79	.	0.38	4.91*	0.42	.	13.66**	14.74***	1.81	0.42	0.42	0.42	4.91*	0.42	0.14	6.22**	6.65**	6.72	0.14	0.14	0.14
LT*Fertilizer	.	1.64	2.00	2.23	.	1.27	19.94***	14.74***	.	1.27	0.40	19.94***	4.69	14.74***	14.74***	19.94***	4.69	0.03	0.03	0.20	0.03	0.07	0.03	0.03
Soil*Fertilizer	6.26*	9.97**	8.31**	17.82***	4.96*	0.02	0.00	0.00	4.96*	0.02	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.07	0.07	0.10	0.07	0.10	0.10	0.10
AC*LT*fertilizer	.	3.07*	3.09*	2.19	.	0.63	0.76	0.00	.	0.63	0.51	0.76	0.00	0.00	0.00	0.76	0.15	0.46	0.46	0.47	0.46	0.46	0.46	0.46
r ²	0.49	0.68	0.73	0.52	0.61	0.63	0.76	0.51	0.61	0.63	0.51	0.76	0.51	0.51	0.51	0.76	0.15	0.46	0.46	0.47	0.46	0.46	0.46	0.46

p*<0.05; *p*<0.01; ****p*<0.001

effect on shoot biomass and total biomasses varied with plant species (significant LT × plants species interaction, *p*<0.01), soil type (significant LT × soil interaction, *p*<0.05) and fertilizer (significant LT × fertilizer interaction, *p*<0.001). Its effects on the shoot/root ratio also varied with fertilizer addition (significant LT × fertilizer interaction, *p*<0.05). Further the presence of both earthworm species affected the shoot/root ratio and this effect varied with plant species (significant AC × LT × plant species interaction, *p*<0.001). Finally we found that the shoot/root and total biomasses were affected by the soil type and fertilizer but plant species differed in their responses (significant soil × plant species and fertilizer × plant species interactions, *p*<0.001). These general results and the significant interactions involving the plant species justify analysing separately for each plant species the effects of the presence of *L. terrestris*, soil type and fertilization (Table 2).

Shoot biomass

L. terrestris increased the shoot biomass of *T. dubium* (+26%) but decreased the shoot biomass of *V. persica* (-24%) (Table 2). Mineral fertilization increased the shoot biomasses of *T. dubium*, *V. persica* and *P. annua* (Table 2) by 24%, 46% and 48% respectively. The shoot biomass of *T. dubium* was 140% higher in the nutrient poor soil than in the nutrient rich one. The significant LT × fertilizer interaction for the shoot biomasses of *V. persica* and *P. annua* (Table 2) showed that effect of LT on these plants species varied with mineral fertilization. LT reduced the shoot biomasses of *V. persica* (-37%) and *P. annua* (-10%) when fertilizer was added (Fig. 1). The significant LT × Soil interaction, (Table 1 and LSmeans comparisons) indicated that the effect of LT on shoot biomass of *T. dubium* varied with soil type. It increased the shoot biomass of the legume (+130%) only in the nutrient rich soil.

Root biomass

Mineral fertilization decreased the root biomasses of *P. annua* (-50%) and *T. dubium* (-25%). The root biomasses of the three plant species were affected by soil type with the highest root biomass in the nutrient rich soil for *P. annua* and *V. persica* while *T. dubium*

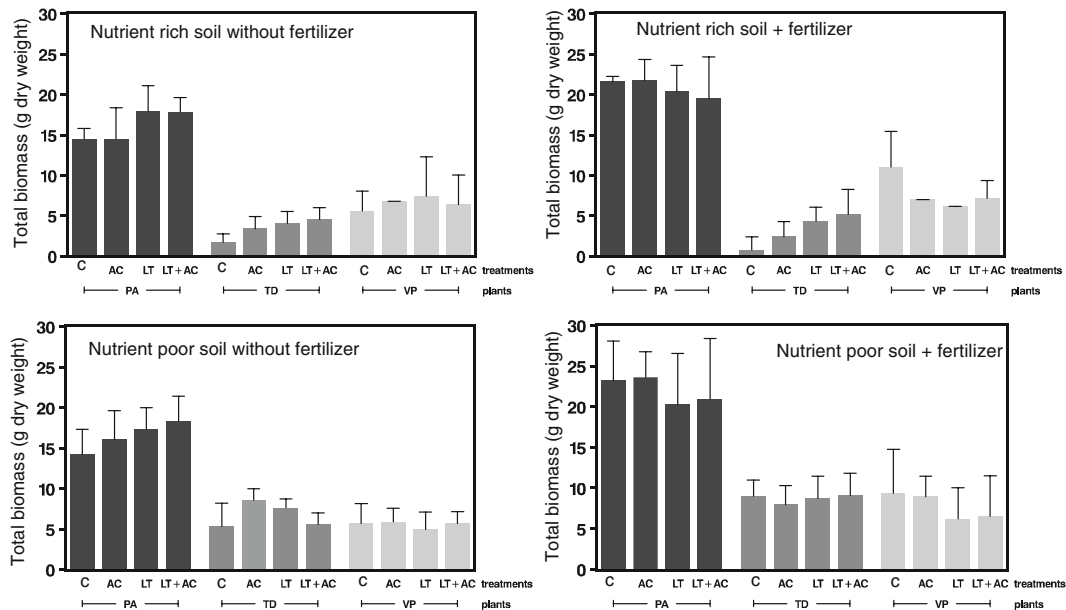


Fig. 1 Effects of earthworms, soil type and fertilizer on the total biomasses of *P. annua* (PA), *T. dubium* (TD) and *V. persica* (VP). Abbreviations: C, control treatment; AC, *A.*

caliginosa only; LT, *L. terrestris* only; LT + AC, combined treatment with *A. caliginosa* and *L. terrestris*

showed higher root biomass in the nutrient poor soil (Table 2).

Total biomass

LT increased by 26% the total biomass of *T. dubium* but decreased the total biomass of *V. persica* by 24%. Mineral fertilization increased the total biomass of *T. dubium* (16%), *P. annua* (31%) and *V. persica* (41%) (Table 2). Significant interactions between LT and fertilizer were found for the total biomasses of *P. annua* and *V. persica* suggesting that effects of LT on these plant species varied with mineral fertilization (Table 2). Without mineral fertilization, LT increased the total biomass of *P. annua* (+18%) while no significant effect of this earthworm was found when the fertilizer was added. For *V. persica* LT reduced (−48%) the total biomass when the fertilizer was added but without mineral fertilization it did not show significant effect (significant LSmeans comparisons and Fig. 1). For *T. dubium* the significant LT × Soil interaction (Table 2) showed that the effect of LT on the total biomass of this plant species varied with soil type. In the nutrient rich soil the presence of LT increased significantly the total biomass of *T. dubium* (+121%) while, no significant effect of LT was found in the nutrient poor soil (Fig. 1). Moreover, in the

nutrient rich soil without earthworms, the mineral fertilization reduced (−55%) the growth of the legume (Fig. 1). The effect of soil type on the total biomass of *T. dubium* varied with fertilizer addition (Table 2, significant Soil × Fertilizer interaction). In the nutrient poor soil, the addition of fertilizer increased (+22%) the total biomass of the legume while no significant effect of fertilizer was found in the nutrient rich soil (Fig. 1).

Shoot/root ratio

The presence of LT increased by 21% the shoot/root ratio of *P. annua*. Shoot/root ratio of *V. persica* was increased (+32%) when both earthworm species were present (Fig. 2 and LSmeans comparisons). Moreover the shoot/root ratios of the three plant species were significantly affected by the soil type and mineral fertilization. Their shoot/root ratios were higher in the nutrient poor soil than in the nutrient rich one (Fig. 2). Mineral fertilization increased the shoot/root ratio of *T. dubium* (193%), *V. persica* (194%) and *P. annua* (335%). Effects of LT on the shoot/root ratio of *P. annua* varied with the soil type (significant LT × soil interaction, Table 2). LT increased (+27%) the shoot/root ratio of *P. annua* only in the nutrient poor soil (Fig. 2 and LSmeans comparisons).

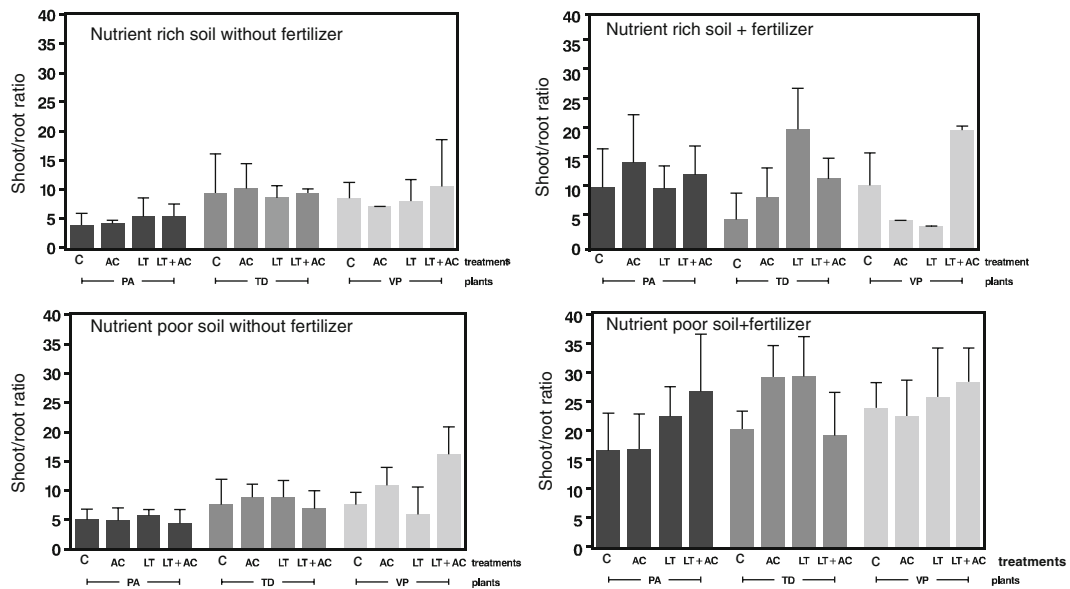


Fig. 2 Effects of earthworms, soil type and fertilizer on the shoot/root ratio of *P. annua* (PA), *T. dubium* (TD) and *V. persica* (VP). Abbreviations: C, control treatment; AC, *A.*

caliginosa only; LT, *L. terrestris* only; LT + AC, combined treatment with *A. caliginosa* and *L. terrestris*

Discussion

We hypothesized that: both earthworm species should affect (1) plant growth and (2) plant resource allocation only in the nutrient and organic matter poor soil; (3) significant interactive effects of the two earthworm species on plant growth and resource allocation should only be found in the nutrient and organic matter poor soil; (4) the impact of *A. caliginosa* and *L. terrestris* on plants should decrease with mineral fertilization.

It is generally thought that positive effects of earthworms on plant growth are more likely in nutrient poor soils than in nutrient rich soils (Brown et al. 2004). Although only *L. terrestris* showed significant effects, our results contradict this hypothesis. Our first hypothesis was not confirmed since LT, increased the shoot biomass and the total biomass of *T. dubium* only in the nutrient rich soil. Our second hypothesis was confirmed since there was no case where an earthworm species changed the shoot/root ratio in the rich soil but not in the poor soil. For example, for *P. annua* the presence of LT increased the shoot/root ratio but only in the poor soil as predicted. Few significant interactions between the anecic and the endogeic earthworms were found in this study, and these interactions did not depend on the soil type.

Consequently, our third hypothesis was not verified. Similarly, our fourth hypothesis was not supported by our results since there was no case where mineral fertilization decreased or suppressed a positive effect of an earthworm species on the growth of a plant species.

Plant growth

Effects of LT on the growth of *T. dubium* varied with the soil type but contrary to our expectation, this earthworm species did not have any significant effects on this plant in the nutrient poor soil. Indeed, LT increased the shoot and total biomasses of *T. dubium* only in the nutrient rich soil and mineral fertilization did not modify these effects (see Fig. 1). These results suggest that LT effect on *T. dubium* was probably not due to an enhancement of mineralization. This hypothesis is supported by the fact that the fertilizer addition had the opposite effect: it increased the total biomass of this plant only in the poor soil. These results should thus be explained by one of the four other mechanisms through which earthworms influence plant growth (Brown et al. 2004; Scheu 2003). A possible mechanism could be the production of growth regulators in the presence of LT (Blouin et al. 2006). Indeed, as microbial biomass is generally higher in clayey soils than in sandy soils (Hendrix

et al. 1998), the activity of LT could have led to a greater production of phytohormones, through the stimulation of bacteria in the nutrient rich soil than in the nutrient poor soil and this resulted in an increase in the growth of *T. dubium*. Moreover, *T. dubium*, as a legume, is less sensitive to an increase in mineralization (Brown et al. 2004; Jenerette and Wu 2004), which further supports our rationale. An alternative explanation for the positive effect of LT on the legume is that this earthworm species could have increased microbial biomass burying organic matter in the soil which could in turn increase the immobilization of mineral nutrient (Van der Heijden et al. 2007) and allowed *T. dubium* to grow better because it is better adapted to nitrogen poor soils. This hypothesis is supported by the fact that the legume had a lower total biomass in the nutrient rich soil than in the nutrient poor one. Finally, the soil type itself did not affect the growth of *P. annua* and *V. persica* suggesting that these plant species are less sensitive to the soil quality than the legume.

Although LT increased the total biomass of *P. annua* without mineral fertilization, it decreased the total biomasses of *P. annua* and *V. persica* when mineral fertilizer was added. This suggests that, in our experiment, without mineral fertilization, LT affected these plants mostly through an increase in mineralization. An explanation for the negative effect of LT on the total biomass of *P. annua* and *V. persica* when mineral fertilizer was added is that this earthworm species could have enhanced the loss of added mineral nutrients through the galleries it produced. This suggests that LT could influence plant growth not only through mineralization of organic matter, at least when the soil contains enough organic matter, but also through its effects on nutrient losses (Barot et al. 2007; Dominguez et al. 2004).

Shoot/root ratio

The effect of earthworms on plant resource allocation is poorly documented and understood (Scheu 2003), but increase of the shoot/root ratio in the presence of earthworms has been documented before (Kreuzer et al. 2004; Scheu et al. 1999). In our study, LT effect on *P. annua* shoot/root ratio varied with the soil type, this earthworm species increased the shoot/root ratio only in the nutrient poor soil. The significant effect of

the interaction between LT and the soil type on the shoot/root ratio is probably due to the strategies plants have evolved to optimize their resource allocation to their root system to efficiently take up nutrients. Indeed, when mineral nutrients are poorly available it is efficient for a plant to increase locally its root biomass in nutrient rich patches, while, if mineral nutrient availability is high enough in the whole soil, it is more efficient for a plant to decrease its total root biomass (Wilson 1988). However, plants have evolved different thresholds, first, in the local nutrient availabilities triggering local root proliferation, second, in the general nutrient availabilities triggering a change in the shoot/root ratio (Hutchings 1988). In this way, the observed variations of the responsiveness to earthworms of the shoot/root ratio with the soil type would reflect the scale and the precision at which root systems exploit the soil (Campbell et al. 1991). Moreover, the possible local release of plant growth factors (Muscolo et al. 1999) in earthworm casts and local changes in soil structure due to earthworm activities are also likely to influence local root densities as well as whole shoot/root ratios. This suggests that studying thoroughly changes in the allocation to the root system and in the root system architecture in the presence of earthworms and in different soils would provide useful information on root foraging strategies.

Interactive effect of both earthworm species on plant growth

Few significant effects of the interaction between LT and AC were found in our study; moreover the observed effects did not change with the soil type. This suggests that mechanisms through which the two earthworm species might influence plant growth do not interact. In our experiment this is also probably due to the fact that few significant effects of AC were found. Contrarily to other experiments using similar biomass of AC (Kreuzer et al. 2004; Wurst et al. 2003, 2005), no significant effect of this earthworm species was found on plant growth in our study. AC individuals were probably less active than those of LT that produced a lot of casts on the soil surface (personal observations). Another complementary explanation is that AC might have enhanced mineralization or triggered the release of phytohormones in the soil without affecting significantly plant growth.

In this case, AC might have influenced the physiology of these plants, for example the allocation of N to the root and aerial system, without affecting their biomass. This hypothesis is supported by the results of another experiment where AC affected the N content of plant without affecting their growth (Laossi et al. 2009). Other studies are needed to test for this kind of effect.

Conclusion

We have confirmed that earthworm effects (especially *L. terrestris*) on plants change with soil type (Doube et al. 1997) and nutrient supply. However, our results suggest that earthworm effects do not necessarily decrease with soil content in mineral nutrients and that earthworms do not affect plant via mineralization only. Consequently, other mechanisms are influential and new experiments are required to predict when and where the different mechanisms are influential. This is required to be able to predict how earthworm effects on plant growth vary with soil characteristics such as the availability of mineral nutrients, organic matter content and texture. The issue should be tackled by systematically comparing earthworm effect on plant, everything else being equal, in many different soils representing various combinations of these characteristics.

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